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# Bilal et al., 2023 (M41) 18-11- 2023

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1 **Reduction in phytotoxicity of a textile wastewater against *Vigna radiata* using *Citrobacter* sp.**  
2 **M41 in a bioaugmented packed bed column bioreactor**

3 **Abstract**

4 The current research was carried out to devise a biological strategy to remediate dyes and metal  
5 ions from textile wastewater. For this purpose, *Citrobacter* sp. M41 was isolated from a textile  
6 wastewater sample. This strain efficiently decolorized > 90% of the added reactive black-5 (RB5)  
7 dye using yeast extract as a C source at 8.5 pH and 30 °C in mineral salt medium containing a  
8 mixture of metals (Pb<sup>2+</sup>, Ni<sup>2+</sup> and Cd<sup>2+</sup>). Moreover, M41 showed >90% decolorization of RB5  
9 beside >80% concurrent removal of hexavalent Chromium [Cr(VI)]. Furthermore, the granulated  
10 corncob and its biochar bioaugmented with M41 and filled within packed bed column based  
11 bioreactors efficiently (>90%) removed the RB5 and Cr(VI) from a textile wastewater. The M41  
12 bioaugmented corncob biochar resulted into a 60.8%, 57.8% and 68.7% decrease in chemical  
13 oxygen demand (COD), electrical conductivity (EC) and total dissolved solids (TDS) of the  
14 wastewater, respectively. A phytotoxicity study conducted by using mung bean (*Vigna radiata*) as  
15 a test crop indicated that the treated synthetic wastewater was relatively less toxic as compared  
16 with the untreated wastewater. Therefore, it can be concluded that *Citrobacter* sp. M41 is an  
17 important bio-resource for treatment of textile effluent polluted with azo dyes and Cr(VI).

18 **Keywords:** *Citrobacter* sp. M41; Wastewater Treatment; Bioaugmented Packed Bed Column  
19 Bioreactors; Hexavalent Chromium; Azo Dyes

20

21 **1 Introduction**

22 Textile sector is reported to be a major producer of colored wastewaters because a significant  
23 amount of the applied dyes (5-50%) remains unsuccessful to attach with the substrate at the time



24 of dying process (Robinson et al., 2001). Apart from dyes, immense amount of some heavy metals  
25 like Cr<sup>+2</sup>, Pb<sup>+2</sup>, Ni<sup>+2</sup>, Cd<sup>+2</sup>, Mn<sup>+2</sup>, Cu<sup>+2</sup>, Co<sup>+2</sup> are also used in dyeing processes and can be found in  
26 textile wastewater (Imran et al., 2015). Among various heavy metals, [Cr(VI)] is excessively  
27 utilized in dyeing processes. It is being used as a part of complexed dyes, oxidant in sulphur and  
28 vat dyeing and a mordant to fix dye color on fabrics during chrome dyeing process (Desai et al.,  
29 2009; Maqbool et al., 2016).

30 Releasing wastewater containing various azo-dyes and heavy-metals into surface water  
31 resources causes various issues including aesthetic problems, restriction in light absorption,  
32 hindrance in movement of oxygen and increase oxygen demands which poses serious threats for  
33 aquatic organisms (Li et al., 2012; Imran et al., 2015, Imran et al., 2019). Literature shows that  
34 several azo dyes and their metabolites as well as Cr(VI) have negative effects on biotic components  
35 of the environment (Mittal et al., 2005; Imran et al., 2019). Azo-dyes and Cr(VI) are found to be  
36 stable in soil and capable of affecting microbial processes and enzymatic activities (Imran et al.,  
37 2015). Hence, it is very important to design some suitable approaches to remove dyes and  
38 hexavalent chromium from textile wastewaters.

39 Recently, various scientists conducted different experiments to show the significance of  
40 microorganisms like bacteria, fungi, actinomycetes and algae for biological treatment of textile  
41 wastewater (Khalid et al., 2012; Hussain et al., 2017; Baig et al., 2019; Imran et al., 2019). Despite  
42 the fact that various fungal species have shown effectiveness in remediation of azo dyes and metal  
43 ions, bacteria are preferred due to rapid removal rate and shorter life span (Elisangela et al., 2009).  
44 Various strains of bacteria from several genera isolated from the wastewaters have shown potential  
45 to decolorize and detoxify various dyes (Hafeez et al., 2018; Imran et al., 2019; Hussain et al.,  
46 2020) and detoxify different metal ions including Cr(VI) (Maqbool et al., 2015). However, few

47 recent studied showed the concurrent elimination of dyes and heavy metal like Cr(VI) by some of  
48 the bacteria isolated from different locations (Anwar et al., 2014; Maqbool et al., 2016; Hussain et  
49 al., 2020; Bilal et al., 2022).

50 Although, few bacterial isolates are already used for concurrent azo dyes decolorization  
51 and Cr(VI) removal, it is much needed to isolate and characterize novel more potential bacterial  
52 isolates for such simultaneous removal and test the possibility of their application for treatment of  
53 the wastewaters. In this regard, the current research work was carried out for isolation of a novel  
54 bacterial strain which might be harboring the capability of contemporaneous removal of reactive  
55 black 5 (RB5) dyes and Cr(VI) in the medium containing the metal ions. The current research is  
56 exclusive because this novel strain has the capability of concurrent removal of dyes and Cr(VI) in  
57 the medium containing other metals. Moreover, this strain was tested for its application for  
58 concurrent decolorization of RB5 and Cr(VI) removal in packed bed column based bioreactors  
59 using the granulated corncob and granulated corncob biochar as support materials for  
60 bioaugmentation with this strain. The untreated and treated textile wastewaters were also tested  
61 for their phytotoxicity using mung bean (*Vigna radiata*) as a test crop.

## 62 **2 Materials and Methods**

### 63 **2.1 Dyes, Chemicals and Culture Media**

64 The major characteristics of the dyes used in this research have been presented in Table S1. A  
65 mineral salt (MS) medium having the composition already reported in Hussain et al. (2013) was  
66 used for isolation of bacteria. However, for providing the metal stress, this medium was added  
67 with different metals ( $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  &  $\text{Ni}^{2+}$  @ 10 mg L<sup>-1</sup> each) in the form of  $\text{CdCl}_2 \cdot 2\text{H}_2\text{O}$ ,  $\text{Pb}(\text{NO}_3)_2$

68 and NiCl<sub>2</sub>. Wherever required, the agar (16 g L<sup>-1</sup>) was added to medium. The composition of any  
69 other medium has been presented wherever used in this study.

## 70 2.2 Collection and characterization of textile wastewater samples

71 The textile wastewater samples were collected from six different industries of Faisalabad  
72 (Pakistan) in pre-sterilized plastic bottles and analyzed for different physicochemical properties  
73 following the standard procedures. The sources and physicochemical properties of effluent  
74 samples have been presented in Table 1. A component of the original samples was preserved (4°C  
75 under dark) which were later on processed to isolate metal tolerant dye decolorizing microbial  
76 strains.

## 77 2.3 Isolation of the potential reactive black 5 (RB5) decolorizing bacterium

78 The bacteria with potential to decolorize RB5 while tolerating the metal ions were isolated  
79 following an enrichment procedure followed by a dilution plating technique as already reported  
80 by Hussain et al. (2013). Among the purified bacterial colonies, 19 fast growing bacterial isolates  
81 were subjected to screening to check their ability for RB5 (200 mg L<sup>-1</sup>) decolorization in MS  
82 containing multi-metal mixture (Cd<sup>2+</sup>, Pb<sup>2+</sup>, Ni<sup>2+</sup>). The isolate M41 was selected for use in further  
83 experiments because it efficiently decolorized RB5.

## 84 2.4 Identification of the Isolate M41

85 Identification of M41 was done by amplifying and sequencing its 16S rRNA. Amplification of the  
86 16S rDNA gene of M41 was done through PCR using 27f and 1492r primers in accordance with  
87 program and reaction mixture previously explained by Maqbool et al. (2016). Purification and  
88 sequencing of 16S ribosomal RNA was done by Macrogen (Seoul, Korea). BlastN analysis was  
89 carried out for comparison of 16S gene sequence of M41 with known sequences in the NCBI  
90 GenBank. Several alignments of sequence of M41 were carried out using clustalX software and a

91 phylogenetic tree was constructed through neighbor joining method (Thompson et al., 1997;  
92 Maqbool et al., 2016).

### 93 2.5 Characterization of *Citrobacter* sp. M41 for its decolorizing potentials

94 The decolorizing capability of *Citrobacter* sp. M41 for selected different azo dyes was  
95 tested in MS media under stress of the mixture of metal ions ( $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Ni}^{2+}$ ). The broth  
96 medium added with respective dye ( $200 \text{ mg L}^{-1}$ ) was inoculated with bacterial culture of strain  
97 M41 to develop a bacterial density ( $\text{OD}_{600}$ ) of 0.2. These samples and the un-inoculated controls  
98 of each dye were kept at  $30^\circ\text{C}$  in static environment. Aliquots of each incubated sample were  
99 taken at different times, centrifuged ( $7000 \text{ rpm}$  for  $10 \text{ min}$ ) and analyzed for estimation of  
100 decolorization using UV-visible spectrophotometer at their specific  $\lambda_{\text{max}}$  (Table S1).

101 Reactive black 5 decolorization by *Citrobacter* sp. M41 was also tested under shaking as  
102 well as static incubations under given conditions. For this purpose, two set of MS broth media  
103 containing RB5 (@  $200 \text{ mg L}^{-1}$ ) were inoculated with strain M41 [optical density ( $\text{OD}_{600}$ ) of 0.2].  
104 All of inoculated tubes and their un-inoculated controls were sealed tightly and one set was  
105 incubated in shaking ( $150 \text{ rpm}$ ) conditions and the other was incubated in static environment at  
106  $30^\circ\text{C}$ . The aliquots were separated and put under centrifugation ( $7000 \text{ rpm}$  for  $10 \text{ min}$ ). In order to  
107 estimate RB5 decolorization, supernatants were assessed using UV-visible spectrophotometer.

108 Glucose, maltose, D-mannitol and yeast extract carbon co-substrates were evaluated to see  
109 the impact on RB5 decolorizing potential of M41. For this, the carbon co-substrates were added in  
110 media separately along with a control without any co-substrate. After inoculation with M41 ( $\text{OD}_{600}$   
111 of 0.2), all the vials as well as their counter un-inoculated controls were kept in static at  $30^\circ\text{C}$ .  
112 During incubation, the samples were taken at different time intervals, centrifuged ( $7000 \text{ rpm}$  for  
113  $10 \text{ min}$ ) and used to estimate decolorization (%) of RB5.

114 Strain M41 was also tested <sup>2</sup> to decolorize RB5 at various pH levels ranging from 5.5 to 9.5  
115 using MS broth media under stress due to the mixture of metal ions. The adjustment of pH of the  
116 media was done using standard solutions of HCl and NaOH. After inoculation with M41 (OD<sub>600</sub>  
117 of 0.2), all the <sup>6</sup> vials along with un-inoculated controls in triplicate were kept in static conditions at  
118 30°C. Over regular time periods, aliquots were collected, centrifuged (7000 rpm for 10 min) and  
119 examined with UV-Visible spectrophotometer for RB5 decolorization.

120 The efficiency of M41 for decolorization of RB5 dye at its varying initial concentrations  
121 was also tested. For this purpose, various concentrations of RB5 (from 25 to 1000 mg L<sup>-1</sup>) were  
122 maintained and inoculated with M41 (OD<sub>600</sub> of 0.2) in the media, separately. All of the inoculated  
123 samples and their un-inoculated controls in triplicates were incubated in static environment at  
124 30°C. Over regular time periods, aliquots of each incubated vial were taken and centrifuged for 10  
125 minutes at 7000 rpm. The supernatants were analyzed to check decolorization of RB5.

126 The strain M41 was also tested to decolorize RB5 in the media containing different levels  
127 of metal ions. MS medium was edited with four separate tertiary combinations (0, 10, 50 and <sup>26</sup> 100  
128 mg L<sup>-1</sup> of each selected metal ion) of the metal ions (Cd<sup>2+</sup>, Pb<sup>2+</sup> and Ni<sup>2+</sup>). Addition of RB5 was  
129 done @ 200 mg L<sup>-1</sup>. Optical density of 0.2 (OD<sub>600</sub>) was obtained after inoculating the strain M41  
130 and incubated in triplicates <sup>2</sup> at 30°C in static incubator. Aliquots were collected, centrifuged (7000  
131 <sup>2</sup> rpm for 10 min) and analyzed for RB5 decolorization.

## 132 2.6. Concurrent RB5 decolorization and Cr(VI) removal by *Citrobacter* sp. M41

133 The strain M41 was also evaluated for concurrent RB5 decolorization <sup>12</sup> and Cr(VI) removal. For  
134 this, 18ml of the medium was added with of <sup>4</sup> Cr(VI) (25 mg L<sup>-1</sup>) using K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> and RB5 (200 mg  
135 L<sup>-1</sup>). The bacterial culture of M41 was inoculated (OD<sub>600</sub> equal to 0.2) and incubated statically in  
136 dark at 30°C. All the experiment was carried in triplicates. Aliquots were taken at specific time

137 intervals. Than it was subjected to centrifugation at 7000 rpm for 10 minutes. Supernatants were  
138 taken and divided in to two portions. One portion was used to measure Cr(VI) reduction by  
139 following diphenylcarbazide (DPC) method as previously explained by Anwar et al. (2014).  
140 However, second portion was used for estimation of RB5 decolorization.

141 <sup>25</sup> 2.7 Application of the Strain M41 For Removal of RB5 and Cr(VI) From a Textile  
142 Wastewater in Packed Bed Columns based Bioreactors

143 2.7.1 Use of Corncob and Biochar As Biosorbent For Concurrent <sup>4</sup> Removal of RB5  
144 and Cr(VI)

145 For this study, corncob biomass obtained from an agricultural field was divided into small pieces  
146 and dried at 60 °C for overnight. After drying, corncob biomass was subjected to grinding using  
147 electrical grinder and sieved through ASTM round stainless sieve of mesh size (5 mm). Granulated  
148 biochar was prepared at experimental area of University of Agriculture Faisalabad using  
149 granulated corncob by method described by Shen et al. (2019). For this purpose, the ground  
150 corncob material was heated with a rate of 10°C/min in a furnace with a limited supply of oxygen  
151 to a maximum temperature of 600°C which was maintained for one hour.

152 After preparation of granulated corncob and biochar, 5% of both biosorbent were added in  
153 separated flasks containing distill water. These flasks were amended with either RB5 (200 mg L<sup>-1</sup>)  
154 or 25 mg L<sup>-1</sup> of Cr (VI) or the combination of both RB5 and Cr(VI). These flasks were given a  
155 shaking incubation (150 rpm) at 30°C in a shaking incubator. All <sup>3</sup> the experiment was conducted in  
156 triplicates including their un-inoculated controls. Aliquots from every sample were collected  
157 followed by centrifugation <sup>3</sup> at 7000 rpm for 10 min. The supernatants than used to analyze  
158 <sup>12</sup> simultaneous treatment of RB5 and Cr (VI) as previously described in above sections.

159 *2.7.2 Use of Granulated Corncob and Biochar Bio-augmented With Citrobacter sp.*  
160 *M41 For Simultaneous Removal of RB5 and Cr(VI) From A Textile Wastewater in*  
161 *Packed Bed Columns based bioreactors*

162 This study also demonstrated the practical use of *Citrobacter sp.* M41 to treat colored textile  
163 wastewater having RB5 dye (200 mg L<sup>-1</sup>) and Cr(VI) (25 mg L<sup>-1</sup>) using the packed bed columns  
164 based bioreactors. The textile wastewater collected from an industrial area of Faisalabad was  
165 colorless and had the pH 7.9 and EC equal to 8.3 dS m<sup>-1</sup>. The centrifuged (10000 rpm for 5 minutes)  
166 wastewater was spiked with RB5 and Cr(VI) up to 200 mg L<sup>-1</sup> and 25 mg L<sup>-1</sup>, respectively. For  
167 this experiment, the granulated corncob and corncob biochar materials were first bioaugmented by  
168 immobilizing the cells of the strain M41 on the surface of these materials by using method of Lou  
169 et al. (2019). To achieve the objective, vertical and cylindrical glass columns (1 feet height and  
170 2.5 cm in diameter) were prepared and separately filled with four different types of materials viz.  
171 granulated corncob, granulated biochar, granulated bioaugmented corncob and granulated  
172 bioaugmented biochar. Low pressure pumps were attached for passing the wastewater having RB5  
173 (200 mg L<sup>-1</sup>) and Cr(VI) (25 mg L<sup>-1</sup>). After every pass of synthetic wastewater from the sorbent  
174 regions, aliquots were taken, centrifuged (7000 rpm for 10 min) and analyzed for simultaneous  
175 removal of RB5 and Cr(VI) as explained in above sections. The supernatants after the final pass  
176 were also analyzed for pH, electrical conductivity, total dissolved solids (TDS) and chemical  
177 oxygen demand (COD) using the standard methods already described in Maqbool et al. (2016).

178 *2.8. Phytotoxicity evaluation of the treated wastewaters*

179 In order to assess the toxicity, the centrifuged (10000 rpm for 5 minutes) untreated as well as  
180 treated wastewaters were separately applied for the germination of mung bean (*Vigna radiata*, cv.  
181 NM-2006) seeds obtained from Ayub Agricultural Research Institute, Faisalabad, Pakistan. Before

182 germination, surface sterilization of the seeds was carried out for ten minutes with 3% hydrogen  
183 peroxide and then seeds were rinsed with deionized water. The triplicate sets of 10 mung bean  
184 seeds in sand in petri plates were irrigated with 10 mL of untreated as well as treated wastewaters  
185 separately. The control was irrigated with equal amount of distilled water. After one week, the data  
186 regarding the germination (%) and length of plumule and radical were noted and statistically  
187 analyzed by analysis of variance followed by Fisher LSD test.

### 188 **3 Results**

#### 189 **3.1 Isolation and Identification of the Strain M41**

190 The results exhibited that the tested 19 bacterial isolates exhibited varying potential for  
191 decolorization of RB5 (Fig. S1). Among these 19 isolates, the isolate M41 exhibited maximum (>  
192 90%) decolorization potential against RB5. Therefore, the isolate M41 was selected for further  
193 experiments. The BlastN analysis of the 16S rDNA gene of M41 indicated that this strain had the  
194 maximum (~ 99%) resemblance with bacterial isolates of genus *Citrobacter*. Additionally, the  
195 phylogenetic tree built by using neighbor joining method also confirmed the position of strain M41  
196 in bacterial strains cluster of genus *Citrobacter* (Fig. 1). Hence, this isolate was named as  
197 *Citrobacter* sp. M41 and the sequence of its 16SrDNA gene was submitted in NCBI GenBank  
198 database under accession No. MT730590. This strain was found to carry out maximum  
199 decolorization of RB5 ( $92.5 \pm 2.4\%$ ) followed by RY2 ( $91.4 \pm 3.0\%$ ) in MS medium added with  
200 metals, but, the decolorization of RR120 and RO16 was found negligible even at last stage of the  
201 incubation (72 h) (Fig. S2).

#### 202 **3.2 Characterization of RB5 Decolorizing Capabilities of *Citrobacter* sp. M41**

203 While studying the effect of shaking and static incubation, *Citrobacter* sp. M41 exhibited higher  
204 RB5 decolorization in static state of incubation in comparison with shaking incubation (Fig. 2a).



205 According to results, 31.3% and 18.1% decolorization of added RB5 was observed under static  
206 and shaking conditions, respectively, within 24 h of incubation. After incubation of 72 h, the  
207 maximum RB5 decolorization was recorded under static incubation (91.4%) followed by shaking  
208 incubation (56.7%).

209 It was observed that addition of carbon co-substrates considerably enhanced the RB5  
210 decolorization by M41 (Fig. 2b). After 24 h, 51.8%, 5.4%, 3.6% and 2.1% of the RB5 was  
211 decolorized while adding yeast extract, maltose, glucose and D-Mannitol, respectively. At the last  
212 stage of the incubation, the maximum RB5 removal was again recorded by adding yeast extract  
213 (85.3%) followed by glucose (37.0%), D-mannitol (19.2%) and maltose (11.9%).

214 Results regarding the effect of varying initial concentrations of RB5 on its decolorization  
215 <sup>9</sup> in the presence of metal ions are shown in Fig. 2c. Results depicted that lower concentrations of  
216 RB5 (25-200 mg L<sup>-1</sup>) showed non-significant impact on RB5 decolorization, however, the higher  
217 initial RB5 concentrations (300-1000 mg L<sup>-1</sup>) exhibited negative effect on RB5 decolorization  
218 (Fig. 2c). Over the incubation period, > 90% of RB5 was decolorized within the media having 25,  
219 <sup>5</sup> 50, 100, 150 and 200 mg L<sup>-1</sup> of initial concentrations of RB5. However, 78.3%, 56.7% and 23.5%  
220 of RB5 was decolorized in the medium containing 300, 500 and 1000 <sup>5</sup> mg L<sup>-1</sup> of initial  
221 concentration of RB5 in the MS media containing the metal ions.

222 Results showed that optimum decolorization of RB5 was achieved at pH 8.5 and 7.5 (Fig.  
223 2d). The maximum decolorization of RB5 over 24 h incubation was recorded at pH 8.5 (48.3 ± 1.5  
224 %) followed by at pH 7.5 (21.8 ± 3.8%) and pH 9.5 (20.3 ± 0.3%). The maximum decolorization  
225 was at alkaline pH 8.5 (80.7 ± 4.8 %) followed by at 7.5 (77.6 ± 1.0 %), 9.5 (37.2 ± 3.5 %), 6.5  
226 (23.5 ± 4.6 %) and 5.5 (8.2 ± 0.2 %) at the end of the incubation.

227 While studying the <sup>1</sup> impact of the varying levels of multi-metal mixture (Cd<sup>2+</sup>, Pb<sup>2+</sup> and  
228 Ni<sup>2+</sup>) on decolorization of RB5 by the strain M41, it was observed that the presence of the multi-  
229 metal mixtures did not inhibit the RB5 decolorization (Table 2). However, they affected the rate  
230 of decolorization of RB5 by the strain M41. Over 72 h incubation, 94.6%, 92.8%, 73.4% and  
231 38.4% decolorization of RB5 was observed in the media containing no metals, lower concentration  
232 of the multi-metal mixtures [Cd<sup>2+</sup> (<sup>16</sup> 10 mg L<sup>-1</sup>) + Pb<sup>2+</sup> (10 mg L<sup>-1</sup>) + Ni<sup>2+</sup> (10 mg L<sup>-1</sup>)], medium  
233 concentration of the multi-metal mixtures [<sup>17</sup> Cd<sup>2+</sup> (50 mg L<sup>-1</sup>) + Pb<sup>2+</sup> (50 mg L<sup>-1</sup>) + Ni<sup>2+</sup> (50 mg L  
234 <sup>15</sup> )] and higher concentration of the multi-metal mixtures [Cd<sup>2+</sup> (100 mg L<sup>-1</sup>) + Pb<sup>2+</sup> (100 mg L<sup>-1</sup>) +  
235 Ni<sup>2+</sup> (100 mg L<sup>-1</sup>)], respectively.

### 236 3.3 Concurrent RB5 and Cr(VI) Removal by *Citrobacter* sp. M41

237 The strain M41 exhibited >90 % removal of added Cr(VI) just in 48 hours (Fig. 3a). Based on  
238 these results, *Citrobacter* sp. M41 was also subjected to see the concurrent removal of RB5 and  
239 Cr(VI) and we have found captivated response as shown in Fig. 3b. Throughout the experiment,  
240 *Citrobacter* sp. M41 decolorize faster RB5 dye than Cr(VI). Over 48 h incubation, *Citrobacter* sp.  
241 M41 exhibited >80% removal of RB5 and Cr(VI) in MS broth containing mixture of metal ions.

### 242 3.4 Application of *Citrobacter* sp. M41 For Treatment of Wastewater Using Bio- 243 augmented Packed Bed Columns based bioreactors

244 Both the granulated corncob and its biochar exhibited considerable potential for removal of RB5  
245 as both of the materials showed >80% decolorization of initially added RB5 dye. Although  
246 maximum removal of RB5 was achieved with biochar (93.4±2.0%) followed by corncob  
247 (83.3±3.4%) during 72 hours (Fig. 4a). But in case of Cr(VI) removal, it was surprising to notice  
248 that corncob exhibited >85% removal of initially added Cr(VI) and biochar exhibited <15%

249 removal of Cr(VI) during 72 hours (Fig. 4b). As long as concurrent removal of RB5 and Cr(VI) is  
250 concerned, granulated corncob showed more considerable results in comparison to granulated  
251 corncob-biochar (Fig. 4c). It was noticed that granulated corncob exhibited 59.6% RB5 removal  
252 along with 63.9% simultaneous removal of Cr(VI). However, granulated corncob-biochar showed  
253 >90% removal of RB5 dye with <10% concurrent removal of Cr(VI) during 72 hours incubation.

254 In bioaugmented packed bed columns based bioreactors, it was observed that <sup>11</sup>removal of  
255 both RB5 and Cr(VI) was gradually increased as times of passing of wastewater from these column  
256 was increased (Fig. 5). After single time passing of the textile wastewater, 24.5%, 43.2%, 35.4%  
257 and 48.1% of RB5 was removed in the columns filled with granulated corncob, granulated corncob  
258 biochar, bioaugmented granulated corncob and bioaugmented granulated corncob biochar,  
259 respectively (Fig. 5a). However, after passing the textile wastewater from the columns for five  
260 times, 82.1, 87.6%, 91.3% and 95.7% <sup>30</sup>of the initially added RB5 was removed in the columns filled  
261 with granulated corncob, granulated corncob biochar, bioaugmented granulated corncob and  
262 bioaugmented granulated corncob biochar, respectively (Fig. 5a). There was no significant  
263 increase in RB5 decolorization after further passing of the wastewaters through the columns. As  
264 for as Cr(VI) removal is concerned, it was noticed that after single time passing of the textile  
265 wastewater, 18.2%, 3.4%, 25.5% and 9.7% <sup>7</sup>of the initially added Cr(VI) was removed in the  
266 columns filled with granulated corncob, granulated corncob biochar, bioaugmented granulated  
267 corncob and bioaugmented granulated corncob biochar, respectively (Fig. 5b). However, after  
268 passing the textile wastewater from the columns for five times, 81.9%, 5.1%, 89.4% and 41.6% of  
269 Cr(VI) was removed in the columns filled with granulated corncob, granulated corncob biochar,  
270 bioaugmented granulated corncob and bioaugmented granulated corncob biochar, respectively  
271 (Fig. 5b). After 10<sup>th</sup> passing, 84.1%, 4.6%, 91.4% and 80.7% of Cr(VI) was removed in the in the

272 wastewater samples treated through columns filled with granulated corncob, granulated corncob  
273 biochar, bioaugmented granulated corncob and bioaugmented granulated corncob biochar,  
274 respectively (Fig. 5b).

275 <sup>5</sup> At the end of the study, when RB5 and CR(VI) were recovered and extracted from the  
276 materials in the columns, it was found that 51.3%, 39.8%, 6.7% and 9.1% of RB5 was recovered  
277 from the materials in columns filled with granulated corncob, granulated corncob biochar,  
278 bioaugmented granulated corncob and bioaugmented granulated corncob biochar, respectively  
279 (Fig. 5c). Similarly, 43.7%, 2.4%, 10.9% and 16.2% <sup>7</sup> of the initially added Cr(VI) was recovered  
280 from the materials columns filled with granulated corncob, granulated corncob biochar,  
281 bioaugmented granulated corncob and bioaugmented granulated corncob biochar, respectively.

282 <sup>10</sup> It is also noteworthy that, at the end of the experiment, a significant decrease in the values  
283 of COD, EC and TDS was observed in the textile wastewater samples treated in different packed  
284 bed columns based bioreactors (Table 3). In this study, 41.9% ( $\pm 3.6$ ), 53.6% ( $\pm 4.3$ ), 63.5% ( $\pm 2.5$ )  
285 and 60.8% ( $\pm 5.1$ ) of COD was observed to be removed in the samples treated through the packed  
286 bed columns filled with granulated corncob, granulated corncob biochar, bioaugmented granulated  
287 corncob and bioaugmented granulated corncob biochar, respectively. The EC values of 3.9, 6.2,  
288 4.1 and 3.5  $\text{dS m}^{-1}$  were observed in the wastewater samples treated through the packed bed  
289 columns filled with granulated corncob, granulated corncob biochar, bioaugmented granulated  
290 corncob and bioaugmented granulated corncob biochar, respectively. However, the EC of the  
291 untreated wastewater was  $\text{dS m}^{-1}$ . Similarly, the TDS values of 513, 981, 532 and 426  $\text{mg L}^{-1}$  were  
292 observed in in the wastewater samples treated through the packed bed columns filled with  
293 granulated corncob, granulated corncob biochar, bioaugmented granulated corncob and

294 bioaugmented granulated corncob biochar, respectively. However, the TDS of the untreated  
295 wastewater was 1365 mg L<sup>-1</sup>.

### 296 3.5 Phytotoxicity of the treated wastewaters against *Vigna radiata*

297 The impacts of untreated wastewater and wastewater treated in different packed bed columns based  
298 bioreactors on different parameters of mung bean has been presented in Table 4. It was observed  
299 that the mung bean seeds showed 93.3% ( $\pm 2.9$ ) and 45.0% ( $\pm 5.0$ ) germination when irrigated with  
300 distilled water and untreated wastewater, respectively. However, the germination values of 75.0%  
301 ( $\pm 5.0$ ), 58.3 ( $\pm 7.6$ ), 91.7% ( $\pm 5.8$ ) and 88.3% ( $\pm 7.6$ ) were observed in the mung bean seeds irrigated  
302 with the wastewater samples treated through the packed bed columns filled with granulated  
303 corncob, granulated corncob biochar, bioaugmented granulated corncob and bioaugmented  
304 granulated corncob biochar, respectively. The plumule lengths of the mung bean seedlings  
305 irrigated with distilled water and untreated wastewater were found to be 12.4( $\pm 0.50$ ) cm and  
306 5.7( $\pm 0.80$ ) cm, respectively. However, the plumule lengths of 9.8( $\pm 0.44$ ) cm, 7.9( $\pm 0.35$ ) cm,  
307 12.9( $\pm 1.1$ ) cm and 12.6( $\pm 1.21$ ) cm were observed in the mung bean seeds irrigated with the  
308 wastewater samples treated through the packed bed columns filled with granulated corncob,  
309 granulated corncob biochar, bioaugmented granulated corncob and bioaugmented granulated  
310 corncob biochar, respectively. Similarly, the radical lengths of the mung bean seedlings irrigated  
311 with distilled water and untreated wastewater were found to be 6.43( $\pm 0.72$ ) cm and 2.77( $\pm 0.31$ )  
312 cm, respectively. However, the plumule lengths of 5.80( $\pm 0.70$ ) cm, 4.07( $\pm 0.57$ ) cm, 6.90( $\pm 0.40$ )  
313 cm and 7.13( $\pm 0.31$ ) cm were observed in the mung bean seeds irrigated with the wastewater  
314 samples treated through the packed bed columns filled with granulated corncob, granulated  
315 corncob biochar, bioaugmented granulated corncob and bioaugmented granulated corncob  
316 biochar, respectively.

#### 317 4 Discussion

318 In this study, total 19 bacterial isolates were observed to have varying capability for RB5  
319 decolorization under stress due to a mixture of metal ions including  $Pb^{2+}$ ,  $Ni^{2+}$  and  $Cd^{2+}$ . Variation  
320 in decolonization can be related to variant adaptation of these isolates for removal of dyes as  
321 already mentioned in previous studies (Maqbool et al., 2016; Baig et al., 2019) or due to a varying  
322 resistance to the metal ions in the medium (Abbas et al., 2016). *Citrobacter* sp. M41 was observed  
323 to harbor the highest potential to decolorize RB5 in mixture of metal ions and was selected for  
324 further studies. Despite that few bacterial strains belonging to genus *Citrobacter* are previously  
325 found having potential for decolorization of azo dyes (Sunkar et al., 2015; Schmidt et al., 2019),  
326 however, the strain M41 is unique because it has the potential not only to decolorize RB5 but also  
327 other related dyes i.e. RY2, RR120 and RO16 even under stress due to metal ions. Moreover, the  
328 strain M41 is also unique because it also harbors the potential for concurrent removal of RB5 and  
329 Cr(VI) while tolerating  $Pb^{2+}$ ,  $Ni^{2+}$  and  $Cd^{2+}$  in the medium.

330 The decolorizing ability of *Citrobacter* sp. M41 was optimized under various incubation  
331 conditions. Results revealed that *Citrobacter* sp. M41 carried out maximum decolorization in static  
332 conditions rather than in shaking conditions. Previous studies have also reported the same (Prasad  
333 and Aikat, 2014; Baig et al., 2019; Bilal et al., 2022). Relatively higher decolorization under static  
334 condition might be due to the fact that it provides partially anoxic environment which more  
335 suitably favors dye decolorization because the transfer of electrons to the azo bond and the activity  
336 of dyes reducing enzymes including azoreductase is assisted by anoxic environment (Tripathi and  
337 Srivastava, 2011; Imran et al., 2015; Imran et al., 2019). Presence of carbon co-substrates also  
338 increased the efficiency of *Citrobacter* sp. M41 to decolorize RB5 in the medium added with  
339 mixture of metal ions. It was found that yeast extract increased the decolorization of RB5 more as

340 compared with other carbon co-substrates. Our results were in line with many other studies who  
341 also found efficient <sup>1</sup> decolorization of different dyes including reactive black-5, reactive red-120  
342 and reactive yellow-2 <sup>4</sup> in the presence of yeast extract as co-substrate (Imran et al., 2016; Baig et  
343 al., 2019; Hussain et al., 2020; Bilal et al., 2022). Higher decolorization of RB5 in the presence of  
344 yeast extract might be related to yeast extract being not only carbon and nitrogen source for the  
345 growth of the microorganisms but also as redox mediator due to the presence of riboflavin in it  
346 which enhances the activity of the decolorizing enzymes including azoreductase (Imran et al.,  
347 2016). Our results showed that optimum decolorization of RB5 by Citrobacter sp. M41 was  
348 obtained at pH 8.5 followed by 7.5. Moreover, a decrease in decolorization was recorded as the  
349 pH moved away from these values. Higher decolorization between 7.5 to 8.5 pH might be  
350 <sup>5</sup> attributed to the fact that the growth of the strain as well as the activity of its decolorizing enzymes  
351 including azoreductase is favored within this range of pH (Johansson et al., 2011). However, it is  
352 needed to have a detailed study about pH effects on dye decolorization by studying the microbial  
353 growth as well as the genes and enzymes which participate in dye removal in the presence of metal  
354 ions. The strain M41 was found to considerably decolorize even the higher concentration up to  
355 1000 mg L<sup>-1</sup> of RB5 but rate was decreased at higher concentrations. This reduction might be due  
356 to microbial toxicity of the dyes or their metabolites at their higher concentration (Das and Mishra,  
357 2016). It is also possible that the active sites of the key enzyme i.e., azoreductase might be  
358 blocked by azo dyes with multiple structures, resulting in reduction of decolorization process.

359 Results showed that at <sup>3</sup> different levels of the mixture of metal ions, M41 strain has a good  
360 potential to decolorize RB5 at lower concentrations, however, decolorization was decreased at  
361 higher concentrations. Previous studies reported the impacts of varying mixtures of metal ions on  
362 decolorization of different dyes (Hussain et al., 2013; Abbas et al., 2016). Interestingly a good

363 potential of RB5 decolorization by M41 was observed even at higher concentrations of the mixture  
364 of metal ions. However, low decolorization of RB5 at higher concentration of the metal ions can  
365 be related to their inhibitory effects, suppression of enzymatic activities as already reported by  
366 (Chen, 2011).

367 The results indicated that *Citrobacter* sp. M41 showed good potential not only for  
368 reduction of Cr(VI) but also for concurrent removal of RB5 dye. Few recent studies have reported  
369 the same findings (Mahmood et al., 2013; Anwar et al., 2014; Maqbool et al., 2016), however, in  
370 this study simultaneous removal was studied under stress due to the mixture of other metal ions  
371 ( $Pb^{2+}$ ,  $Ni^{2+}$  and  $Cd^{2+}$ ). Although mechanism of Cr(VI) removal by this strain has not been explored,  
372 however, few previous studies described that chromate play important role as electron acceptor in  
373 order to gain energy for Cr(VI) removal (Maqbool et al., 2015). Nevertheless, more investigations  
374 and research should be carried out to reveal a precise mechanism for Cr(VI) reduction, but, it is  
375 noteworthy that *Citrobacter* sp. M41 exhibited a good potential for concurrent RB5 decolorization  
376 and Cr(VI) removal which make it worthwhile bio-resource for its possible application in  
377 biotreatment of textile wastewater.

378 In the present study, practical application of *Citrobacter* sp. M41 was also carried out in  
379 packed bed columns based bioreactors for treatment of a textile wastewater containing RB5 (200  
380  $mg L^{-1}$ ) and Cr(VI) ( $25 mg L^{-1}$ ). Relatively low sorption of the Cr(VI) on the corncob biochar  
381 might be due to the fact that the surface of this biochar mostly contains the negative charges due  
382 to the presence of a high magnitude of hydroxyl groups at alkaline pH (Tan et al., 2020). This  
383 study indicated that the corncob granules might serve as good materials for biosorption of Cr(VI)  
384 as well as RB5 from the solutions concurrently. While studying the treatment of textile wastewater  
385 containing RB5 (200  $mg L^{-1}$ ) and Cr(VI) ( $25 mg L^{-1}$ ) in packed bed columns based bioreactors, it



386 was observed that both <sup>4</sup> RB5 and Cr(VI) were significantly removed in the columns containing  
387 granulated corncob biosorbent as well its biochar (Figure 5). However, significant amount of RB5  
388 was removed in the columns containing corncob biochar but Cr(VI) was not removed in these  
389 columns. However, <sup>13</sup> it is noteworthy that a significantly higher removal of Cr(VI) was observed in  
390 case of the column filled with bioaugmented biochar. Similar type of <sup>11</sup> removal of RB5 and Cr(VI)  
391 was also observed in a biochar packed bioreactor in which the biochar was bioaugmented with a  
392 dyes decolorizing strain *Pseudomonas putida* (Shahid et al., 2015). It is quite possible that the  
393 immobilization of M41 on the corncob biochar might have changed some surface properties of the  
394 biochar leading towards sorption and removal of the Cr(VI). However, there is need to get a deep  
395 insight into this process by studying the surface properties of the biochar. Further investigations  
396 revealed that bio-augmentation of corncob and corncob biochar with *Citrobacter* sp. M41  
397 improved RB5 and Cr(VI) removal. The corncob and/or its biochar bioaugmented with  
398 immobilized functional microbial strains serve as a suitable combination for remediation of the  
399 wastewaters starting from the sorption on the solid surfaces followed by the decolorization of the  
400 dye molecules <sup>13</sup> and reduction of the Cr(VI) by the functional bacteria on the surface interface of  
401 the solid material. The decolorization of RB5 <sup>13</sup> and reduction of Cr(VI) by the immobilized strain  
402 M41 was supported by the fact that a relatively lower amount of either RB5 or Cr(VI) was  
403 recovered from the bioaugmented corncob and its biochar materials as compared to the corncob  
404 and its biochar without any microbial immobilization (Figure 5C). Such sorption and then  
405 degradation/reduction has to be found in some previous studies focused on removal of different  
406 pollutants including dyes and metals from the wastewater using the functional microbial cultures  
407 immobilized on biochar materials (<sup>24</sup> Shahid et al., 2015; Abu Talha et al., 2017; Bharti et al., 2019;  
408 Ayed et al., 2021). In this study, COD, EC and TDS were also significantly decreased in the

409 wastewaters treated in the packed bed column bioreactors containing the bioaugmented corncob  
410 or its biochar. It further supports the treatment of the textile wastewaters in the bioreactors. While  
411 studying the treatments of a textile wastewater, Ceretta et al. (2017) reported that COD of the  
412 wastewater was significantly decreased due to a bacterial consortium in the presence of carbon co-  
413 substrates.

414 The phytotoxicity study conducted with *Vigna radiata* (mung bean) seeds while  
415 considering the germination rate and length of the plumule and radical in the mung bean seedlings  
416 indicated that the toxicity of all the treated textile wastewaters was significantly decreased as  
417 compared with the untreated textile wastewater (Table 5). Although several studies have been  
418 performed for the treatment of textile wastewaters, however, there are relatively few studies which  
419 also focus on studying the phytotoxicity of the treated wastewaters (Najme et al., 2015; Ceretta et  
420 al., 2017; Ayed et al., 2021). Our finding is in accordance with these few studies who reported the  
421 reduction in phytotoxicity of the dye solutions or dyes loaded textile wastewaters after treatment  
422 by exploiting the bacterial strains (Najme et al., 2015; Ceretta et al., 2017). For example, Ceretta  
423 et al. (2017) observed that the toxicity of a textile wastewater taken in terms of impacts on  
424 germination and radical length of *Lactuca sativa* seeds was significantly decreased after its  
425 treatment by a bacterial consortium. Similarly, Najme et al. (2015) reported that the impacts of a  
426 reactive yellow-2 loaded synthetic wastewater on germination, plumule length and radical length  
427 of mung bean were significantly reduced when it was treated with a bacterial strain *Serratia* sp.  
428 RN34. Recently, Ayed et al. (2021) reported that the phytotoxicity of a textile wastewater against  
429 the *Triticum durum* L. and *Cucurbita pepo* L. seeds was significantly reduced after its treatment  
430 through a consortium composed of bacterial and micro-algal strains. The reduction in toxicity of  
431 the textile wastewater might be not only due to reduction in load of Cr(VI) and RB5 in the treated

432 textile wastewater due to biosorption in the columns but also due to biotransformation of these  
433 contaminants into relatively less toxic forms. For example, Cr(VI) has been reported to be reduced  
434 into less toxic form by several dyes decolorizing bacterial strains including *Pseudomonas putida*,  
435 *Serratia proteamaculans*, *Pseudomonas aeruginosa* and *Acinetobacter junii* in a number of  
436 previous studies (Mahmood et al., 2013; Anwar et al., 2014; Maqbool et al., 2016; Hussain et al.,  
437 2020). Hence this study shows that the application of the dye decolorizing strain *Citrobacter* sp.  
438 M41 as a bioresources in bioaugmented biochar based packed bed column bioreactors might lead  
439 towards the treatment of textile wastewaters along with reduction in their phytotoxic effects.

## 440 **Conclusions and Future Perspectives**

441 It is concluded that *Citrobacter* sp. M41 isolated from textile wastewater possesses good potential  
442 for concurrent removal of RB5 and Cr(VI) even in the medium under stress due to a mixture of  
443 metal ions. Moreover, enhanced concurrent removal of RB5 and Cr(VI) in corncob packed bed  
444 column bioreactor supported with bio-augmentation of *Citrobacter* sp. M41 make it an important  
445 bio-resource to devise technologies for simultaneous RB5 decolorization and Cr(VI) removal  
446 using corncob as sorbent material in column bioreactor. However, there is still need to further have  
447 a thorough understanding of processes involved in concurrent removal of azo dyes and Cr(VI) in  
448 bench scale bioaugmented column bioreactors not only by studying the surface properties of the  
449 materials but also by targeting the metabolites of the dyes.

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