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**Submission date:** 26-Jan-2022 04:51AM (UTC+0500)

**Submission ID:** 1748191020

**File name:** Clean\_MS.docx (3.16M)

**Word count:** 8312

**Character count:** 47754

# A methodological framework integrating habitat suitability and landscape connectivity to identify optimal regions for insecticide application: a case study in Tongzhou, China

## 1 Abstract

2 Insecticide application is still a standard method in pest management. However, the current regions for  
3 insecticide application usually focus on the host of the pest but ignore the migration pathways,  
4 continually resulting in pests' periodic outbreaks. This research provides a valuable modeling  
5 framework to identify optimal regions for insecticide application to solve this problem. This modeling  
6 framework optimizes insecticide areas based on habitat suitability and landscape connectivity,  
7 combining the MaxEnt and circuit theory. We conducted a case study in Beijing, China, to elucidate  
8 the application of this modeling framework. The input data included 517 occurrence points of  
9 *Semiothisa cinerearia* collected in the fieldwork and the environmental variables related to the eco-  
10 physiological characteristics of this species. The results indicated that the Chinese scholar tree  
11 (*Styphnolobium japonicum*) was the most vulnerable species. Increasing the pesticide region by 11.9%  
12 at some crucial corridors of the target species, the pest management effectiveness increased by 27.7%,  
13 compared with the pesticide applied only to the Chinese scholar trees. In summary, this study proposed  
14 a convenient and efficient modeling methodology for planning optimal regions for insecticide  
15 application.

16 **Keywords:** *Styphnolobium japonicum*, Pest control, Ecology model, Species distribution models,  
17 Circuit theory, Invasive species management

## 18 1 Introduction

19 Pest outbreaks in natural and artificial forest ecosystems result in degradation of tree fitness, thus  
20 significantly increasing rates of tree mortality (Ferrenberg, 2016; Marini et al., 2021). For example,  
21 *Semiothisa cinerearia* (Bremer et Grey; Lepidoptera: Geometridae) has caused severe damage to the  
22 Chinese scholar tree (*Styphnolobium japonicum*), a widely planted urban green tree species in China  
23 and East Asia (Tang et al., 2014; X.-Y. Zhu et al., 2020). The *S. cinerearia* consumes the entire tree  
24 leave in 2-3 days and causes disease and ultimately death to the host tree (Wu et al., 1998; Zhang et  
25 al., 2013). In some cities, 50% to 100% of *S. japonicum* had been hosted by *S. cinerearia* (Zhang et  
26 al., 2013) and therefore warrant special attention to control urban green trees loss, especially in like  
27 Beijing where air pollution is a severe issue. Additionally, over the last decades, climate change has  
28 triggered an increase in the range expansion and outbreaks of some pests (Fernandez-Carrillo et al.,  
29 2020; Venäläinen et al., 2020). Forest insect pace with climate change are expected to be highly  
30 propagative, due to shorter life cycle, and have higher fecundity and survival rate (Jactel et al., 2019).  
31 Spraying insecticide is still a common practice in pest management, especially in the urban modified  
32 forestry ecosystems (Meissle et al., 2010).

33 However, using insecticides could not wholly solve some pests' periodic/sporadic outbreaks. For  
34 example, *Simulium chatteri* (a blackfly) broke out periodically in South Africa, even though some wide  
35 insecticide control measures have been conducted (Rivers-Moore et al., 2008). Instead, unchecked and  
36 excessive insecticides caused damage to the natural ecosystem and would be lethal for the pests' natural  
37 predators (Frolov, 2015; Tang et al., 2008). The ever-growing list of pests is damaging natural and  
38 artificial forest ecosystems worldwide (Aukema et al., 2010; Loo, 2008; Ramsfield et al., 2016; Santini

39 et al., 2013). Therefore, it is urgent to determine the efficient regions for pesticides to decrease the risk  
40 of periodic outbreaks and the damage of insecticides to the natural ecosystem.

41 The hypothesis that some insects occupy multiple niches (or hosts) during their life histories and  
42 migration processes could be the basis to explain these issues (Buckley et al., 2010; Melin et al., 2007;  
43 Singer, 1983). For instance, the Asian long-horned beetle (*Anoplophora glabripennis*) attack more than  
44 15 families of tree species throughout their invasion pathways (Dodds and Orwig, 2011; Haack et al.,  
45 2010). Therefore, if the controlling measures had been conducted at the most suitable habitat (e.g., the  
46 food source of the pest), the other adjacent and suitable niches would still provide habitat to the  
47 survived individuals of the pests (Cowley et al., 2015; Demirozer et al., 2012; Dupas et al., 2014; Laćan  
48 and McBride, 2008). This hypothesis implied that if the spread pathways had not been cut off, the  
49 periodic outbreaks would possibly happen again (Tang et al., 2008). Therefore, not only the habitat  
50 suitability should be considered (Phillips et al., 2006), but the dispersal pathways should also be cut  
51 off by anthropogenic barriers to decrease the risk of secondary pest outbreaks (Dickson et al., 2019;  
52 Hlásny et al., 2016).

53 Previous studies (Marchioro and Krechmer, 2021; Reil et al., 2018; Schöneberg et al., 2020) merely  
54 estimated the habitat suitability or landscape connectivity to optimize the pest management regions.  
55 Habitat suitability and landscape connectivity are two critical requirements for periodic pest outbreaks  
56 (Zhu et al., 2020). Still, little has been done to integrate these two essential requirements to identify  
57 regions for efficiently applying insecticide for pest management (Elmes et al., 2019). On the one hand,  
58 habitat suitability quantifies the suitable niche of the targeted pest. Thus the realized niche of this  
59 targeted species (specific environmental condition) would identify (Zhu et al., 2020). The MaxEnt  
60 model was a commonly used species distribution model (SDM) to estimate the habitat suitability  
61 (Phillips et al., 2017; Phillips and Dudík, 2008; Yang et al., 2018), especially for pest management  
62 (Kumar et al., 2023; Wang et al., 2010), which is based on a machine-learning algorithm to predict the  
63 suitable habitat distribution with the maximum entropy solution (Phillips et al., 2006). On the other  
64 hand, landscape connectivity measures the dispersal ability in the pest invasion pathways (Dickson et  
65 al., 2019). As a result, the potentially spatial movement patterns across their incontinent and  
66 heterogeneous habitats can be simulated (Dupas et al., 2014; Viani et al., 2018). The circuit theory  
67 estimates the landscape connectivity of species based on the equivalence between graph theory  
68 representations of random flow of electrons in a circuit (McRae et al., 2008). Many studies applied this  
69 method on species invasion (Elmes et al., 2019; Falaschi et al., 2018), distribution range shift of  
70 endangered mammals (Puddu and Maiorano, 2016; Zhao et al., 2019), gene flow of a disease vector  
71 (Medley et al., 2015), and phylogeographic analysis (Dupas et al., 2014). Therefore, it is necessary to  
72 develop an integrated approach considering both habitat suitability and landscape connectivity,  
73 contributing to the prevention of periodic pest outbreaks.

74 This research provided a useful modeling framework integrating MaxEnt and the circuit theory to  
75 identify optimal regions for insecticide application of *S. cinerearia*, as a case study in Tongzhou  
76 district, Beijing, China. The result was a raster map representing the SDM-circuit values. This article  
77 aimed at 1) constructing a combine model, integrating MaxEnt and circuit theory using the field  
78 surveyed data points; 2) testing the hypothesis that *S. cinerearia* occupied multiple niches in an urban  
79 ecosystem, and 3) customizing the strategy for efficient and economical pest control in their inherent  
80 outbreak regions and dispersal pathways.

## 81 **2 Materials and Methods**

### 82 **2.1 Study area**

83 The study area was the entire region of Tongzhou District (39°36'34"02"N, 116°32'-116°56'E, area  
84 912.34 km<sup>2</sup>, altitude 8.2-27.6 m) of Beijing, China (Figure 1). The mean annual temperature record in  
85 the area is 11.3°C, while the average annual precipitation is 620 mm, and the frost-free period is about  
86 190 days per year. The outbreaks of *S. cinerearia* in 2013 had critically damaged the urban green trees  
87 in the study area (Tang et al., 2014). Some insecticides were subsequently adopted for pest control, but  
88 based on our fieldwork findings, the population of *S. cinerearia* persisted with periodic outbreaks in  
89 the area (Figure 1). The adverse impact in this sprawling city caused by the *S. cinerearia* was a heavy  
90 financial burden to the local government. A new city is in the construction and development phase  
91 here, as the Chinese capital sub-center (Wanghe et al., 2019). Therefore, it is essential to efficiently  
92 control pest outbreaks in the future, which is critical to the new city's development.

## 93 2.2 Sampling

94 We randomly set 24 quadrats (3km × 3km) in the urban and rural areas of Tongzhou District (Figure  
95 1). To achieve random placement of those quadrats, we first divided the study areas into 101 grids.  
96 Secondly, 1,000 random points were generated on the grids. These random points were developed by  
97 the "Create Random Points" tool in the system toolbox of ArcGIS 10.2.2 software. After repeating the  
98 second step 10 times, the top 24 quadrats with the most randomly generated points were selected as  
99 the sampling quadrats. Fieldwork was conducted from June to August 2018. We searched and collected  
100 the target specimens *S. cinerearia* using the "Sweep Sampling Method" (Gadagkar et al., 1990) in each  
101 quadrat, in which we systematically swept all the ground level vegetation (< 2.5m). The occurrence  
102 points were recorded with the help of a Global Position System (GPS) with the World Geodetic System  
103 1984 (WGS 84) coordinate system. Five hundred seventeen individuals of *S. cinerearia* were collected  
104 in the fieldwork (Figure 1 and Table S1). The collected voucher specimens were deposited in the  
105 Natural Museum of Beijing Forestry University, Beijing, China.

## 106 2.3 The scope, assumption, and framework of the modeling

107 Figure 2 illustrates the modeling framework of this study. A combine model, integrating MaxEnt and  
108 the circuit theory was applied. The scope of this combination was to predict the dispersal ability of the  
109 target pest relating to the original habitat suitability without any control measurements and the changes  
110 of habitat suitability after some control measures were conducted. This study was based on certain  
111 assumptions. First, we assumed that the results of MaxEnt represent the present habitat suitability  
112 before pest control. Second, we assumed that the control measures change the current habitat suitability  
113 of the pest. By comparing the SDM-circuit values among those different scenarios (explained in  
114 Section 2.8), the pest management effectiveness among the scenarios was identified.

## 115 2.4 Environmental variables

116 Eight environmental variables (Table 1) were selected to model habitat suitability based on current  
117 knowledge of the eco-physiological characteristics of targeted species (Jiang et al., 2018; Lozier and  
118 Mills, 2011), especially in the urban areas (Godefroid et al., 2020). Additionally, our study area  
119 landscape is plain and relatively small, without apparent altitude and climate differences, so we did not  
120 select altitude and climate as the environmental variables.

121 The environmental variables were extracted from the forest resource inventory database and the land-  
122 use dataset (Table 1). The forest resource inventory database in 2015 was a polygon shapefile provided  
123 by the Beijing Tongzhou Forestry Bureau. The locations of *S. japonicum* (the primary host of *S.*  
124 *cinerearia*), the percentage coverages of forests, and shrubs and meadows in a cell were extracted from  
125 this database as raster maps at the resolution of 10m × 10m. The land-use dataset in 2015 was



126 interpreted from a series of Gaofen-2 satellite images (resolution = 0.80 m) (Wanghe et al., 2020b,  
127 2020a, 2019). Subsequently, four environmental variables, the distance from a raster to the nearest  
128 building, road, cropland, and water, were derived from the land-use dataset using the Euclidean  
129 Distance System Toolboxes of ArcGIS 10.2 software (Environmental Systems Research Institute Inc.,  
130 Redlands, USA).

## 131 2.5 Species distribution model (SDM)

### 132 2.5.1 Correlation analyses of the environmental variables

133 To avoid model overfitting in calibration caused by multicollinearity among predictors (Warren and  
134 Seifert, 2011), we conducted the pairwise correlation ( $r$ ) analyses among the environmental variables  
135 by the Band Collection Statistics Tool of ArcGIS 10.2 software, Following Dormann *et al.*, (2013)  
136 suggestion (conservative threshold of  $r = 0.7$ ) All eight environmental variables exhibited low  
137 bivariate correlation coefficients  $< 0.7$  (Table S2) in the current study. Therefore, all eight variables  
138 were retained for further analyses.

### 139 2.5.2 Parameter setting of MaxEnt model

140 The MaxEnt model version 4.4.1 (Phillips et al., 2006) was used to estimate the habitat suitability from  
141 the 517 surveyed occurrence points of *S. cinerearia* and the eight derived environmental variables of  
142 these occurrence points. The model was submitted to 10 repetitions, and for each repetition, cross-  
143 validation was performed. The occurrence dataset was randomly split into ten equal parts; each was  
144 employed as the model built with the remnants nine test dataset. Accordingly, all records were  
145 practiced nine times for training and one for testing the model, and all the observations were equally  
146 weighted in the training and testing of the model (Merow et al., 2013). A convergence threshold of  
147 0.00001 was inputted, with 500 interactions and 10,000 background points. The Jack-Knife test  
148 quantified the participation of each environmental variable in the modeling. To avoid overfitting,  
149 MaxEnt was operated, allowing only "linear", "product" and "quadratic" features (Falaschi et al., 2018).  
150 The accuracy of the final result was assessed by the area under the receiver operating characteristics  
151 curve (AUC). The range of AUC was from 0 to 1, and the relatively high value of AUC indicated that  
152 the model's prediction result was relatively accurate and agreeable (Figure S1) (Phillips et al., 2006).  
153 This approach produced a raster habitat suitability map with values ranging from 0 to 1.

## 154 2.6 Connectivity model

155 The connectivity model was built using Circuitscape version 4.0.5 (McRae and Shah, 2014), a software  
156 designed to model habitat connectivity using the electrical circuit theory (McRae et al., 2008). The  
157 input was a conductance map from the habitat suitability map estimated by the MaxEnt model. The  
158 standard conductance of each cell was defined by rescaling the habitat suitability map between 0  
159 (minimum suitability) to 1 (maximum suitability). Circuitscape ran in advanced mode to allow for the  
160 use of arbitrary current sources and grounds. The cells at the host location were set as current sources  
161 of one ampere. The cells at the perimeter of the study areas represented the ground of one ohm,  
162 following Falaschi *et al.*, (2018). Thus, the model could predict the possible spread of *S. cinerearia*,  
163 considering all possible directions in one fell swoop (Cowley et al., 2015). The current densities of the  
164 cells estimated by this connectivity model were defined as the SDM-circuit values. The higher SDM-  
165 circuit indicated cells with more necessary net passage probabilities for random movement, moving  
166 from one raster to another (McRae et al., 2008). The movement boundaries and the "pinch points"  
167 among populations were identified by Kernel Density analyses (Silverman, 2018) based on the SDM-  
168 circuit values, using the Kernel Density Tool (Silverman, 2018) of ArcMap 10.2 software (ESRI,  
169 2011).

## 170 2.7 Scenario developments

171 The SDM and connectivity model identified the regions of suitable habitat and corridors of *S.*  
172 *cinerearia*. However, widely spraying insecticide at all those regions was costly and would further  
173 impose some unfavorable impacts on the ecosystem, such as damaging the pests' natural enemies  
174 Therefore, a cost-effective solution was to detect the significant and economical regions/areas for  
175 chemical control. To determine that cost-effective solution, we developed five scenarios and compared  
176 the pest management effectiveness of those scenarios.

177 Firstly, we defined that part A was the host's location and part B was the location of other tree species  
178 except for <sup>36</sup> host. Secondly, part B was divided into five sub-parts based on the SDM-circuit values  
179 ranked by Jenks natural breaks method (Chen et al., 2013): top 20% (sub-part B20), top 40% (sub-part  
180 B40), top 60% (sub-part B60), top 80% (sub-part B80), and top 100% (sub-part B100). Those natural  
181 break classifications were in terms of natural groupings inherent to their SDM-circuit values, whose  
182 class breaks were identified that the best group was similarly valued and that maximizes the differences  
183 between classes (ESRI, 2011). Thirdly, the regions for insecticides of scenario  $B_n$  were defined as the  
184 union of part A and sub-part  $B_n$  ( $n = 20, 40, 60, 80,$  and  $100,$  respectively) (Figure S2). Based on the  
185 second and third assumptions described in section 2.3, the habitat suitability of the regions for  
186 insecticides was modified to a value approaching zero. The habitat suitability of the regions without  
187 insecticides remained unchanged. Finally, the five scenarios were independently inputted to the  
188 connectivity model.

## 189 2.8 Scenario comparisons

190 To examine pest management performance/effectiveness, those five scenarios were compared with the  
191 connectivity model's original result (without the assumptive scenario). The "fuzzy numerical"  
192 algorithm (Jiang and Chen, 2003) (Equations 1 and 2) was implemented, allowing a cell-to-cell  
193 comparison between two raster maps. In Equations 1 and 2, map A was the initial result of the  
194 connectivity model, and map B was one of the scenarios. The raster maps A and B were in the same  
195 location with the same rows and columns. The  $a_i$  and  $b_i$  were one cell on the maps A and B in the same  
196 location, and  $j$  was the total number of cells. Consequently,  $S_{(a_i, b_i)}$  depicted a new map highlighting the  
197 divergences between maps A and B.  $S_{(A, B)}$  referred to the average differentiation between the entire  
198 maps of A and B, ranging from zero to one: one meaning two completely different maps, and zero  
199 representing an identical pair. Therefore, a higher  $S_{(A, B)}$  indicated a better pest management  
200 performance of a sub-scenario than the original result (Visser and De Nijs, 2006). Those calculations  
201 were programmed by Python 2.2 using the ArcPy site package (ESRI, 2011).

202

$$S_{(a_i, b_i)} = 1 - \frac{|a_i - b_i|}{\max(|a_i|, |b_i|)} \quad \text{Equation 1}$$

$$S_{(A, B)} = 1 - \frac{\sum_{a_i=b_i=1}^j S_{(a_i, b_i)}}{j} \quad \text{Equation 2}$$

203 Ultimately, a regression model between the two variables,  $S_{(A, B)}$  and the control area of sub-part  $B_n$ ,  
204 was conducted by the basic Trendline R package. The computer device executing all analyses was a  
205 Mac Pro (2019) graphic workstation with 3.5GHz Intel Xeon W CPU, 32 GB RAM, and Windows 10  
206 Professional operating system.

## 207 3 Results

### 208 3.1 *S. cinerearia* habitat suitability

209 The AUC value ( $0.895 \pm 0.018$ , mean  $\pm$  standard deviation) indicated that the results of SDM had a  
210 receivable result (Figure S1). The high AUC value (compared to the possible maximum 1.0) indicated  
211 that the quality of habitat suitability map derived by the MaxEnt models was high and the sampling  
212 quadrats were sufficient (Phillips et al., 2006). The relative importance of environmental variables was  
213 assessed by the Jack-Knife test (Figure 3), which implied that forests and hosts appeared to have the  
214 most useful statistical information. The contributions of environmental variables are in Table S3. The  
215 habitat suitability map of *S. cinerearia* (Figure 4) illustrated that the average suitability in the entire  
216 study was 0.1231 (SD = 0.1791). Comparably, the mean suitability in the location with *S. japonicum*  
217 was 0.5315 (SD = 0.1468), while the average suitability in the regions without *S. japonicum* was 0.1101  
218 (SD = 0.1640). Our results indicated that the places of *S. japonicum* were ordinarily more suitable for  
219 *S. cinerearia* than other land-use types.

220 We compared the habitat suitability among the top ten widely planted tree species in Tongzhou  
221 District (Figure 5). *S. japonicum*, the host of *S. cinerearia*, provided the most suitable habitat for this  
222 pest insect. In other words, *S. japonicum* was the most vulnerable species to *S. cinerearia*, compared  
223 with the other nine tree species, which was also widely planted in the study area. The genus *Populus*  
224 (*P. tomentosa* and *P. canadensis*) presented the least suitable habitat to *S. cinerearia* compared to the  
225 other eight species.

### 226 3.2 *S. cinerearia* habitat connectivity

227 The map of current densities/SDM-circuit values (Figure 6) indicated the passage probabilities of the  
228 target pest. A cell with a high current density value implied that it would provide great landscape  
229 connectivity to *S. cinerearia* (McRae et al., 2008). Our results showed that the mean current density of  
230 the host tree species (*S. japonicum*) was 39.268 (SD = 30.456), while the average current density of  
231 other tree species was 25.973 (SD = 25.892).

232 Six populations with high current densities were distinguished (Figure 6). Our results recognized  
233 some “pinch points” (Figure 6). The “pinch points” were narrow linear pathways of *S. cinerearia*  
234 between or within those populations.

### 235 3.3 Scenarios comparisons

236 The area of part A is 27.03km<sup>2</sup>, and the areas of sub-parts B20, B40, B60, B80, and B100 (the red  
237 patches in Figure S2) are 3.22, 20.55, 65.47, 137.90, and 253.28km<sup>2</sup>, respectively. Figure 7 describes  
238 the differences in management performances among these scenarios. While the effectiveness of pest  
239 management on the host was similar from scenarios B20 to B100, by increasing the area of chemical  
240 control, the passage probabilities of *S. cinerearia* in the other tree species were decreasing.

241 Figure 8 displays the regression model results between two variables,  $S_{(A, B)}$  and  $A$  (the area of the  
242 sub-part  $B_n$ ). The most suitable model was the polynomial regression, whose  $R^2 = 0.9937$ . This result  
243 indicated that management effectiveness increased while enlarging the regions of chemical control  
244 when  $A < 200$  km<sup>2</sup>. The marginal benefit of scenario B20 was the maximum among all scenarios.  
245 Therefore, scenario B20 was the most efficient one to conduct chemical prevention, in which the areas  
246 of chemical application were increased only by 3.22 km<sup>2</sup> (11.9%). At the same time, the connectivity  
247 of the other trees species was decreased by 27.7%.

## 248 4 Discussion

### 249 4.1 The implications for pest management



250 The list of forests being damaged by pests is growing worldwide (Aukema et al., 2010; Loo, 2008;  
251 Ramsfield et al., 2016; Santini et al., 2013). Compared with natural forests, urban trees are more  
252 vulnerable to pest incursions (Paap et al., 2017). The limited tree diversity in the urban ecosystem  
253 further predisposes urban forests to potentially devastating pest outbreaks (Laćan and McBride, 2008;  
254 Paap et al., 2017). For example, ash is a widely planted city street tree in the United States, consisting  
255 primarily of a limited number of cultivars of white and green ash (*Fraxinus americana* and *Fraxinus*  
256 *pennsylvanica*) (MacFarlane and Meyer, 2005). This situation resulted in low genetic diversity and  
257 enhanced the risk of pest incursions to urban plants. As a result, extensive damage occurred in urban  
258 greening of the northeastern United States with the accidental incursions of the emerald ash borer  
259 (*Agrilus planipennis*) (Kovacs et al., 2010) In the current study, the urban forests were composed of  
260 dozens of different tree species (Figure 5). However, the periodic outbreaks of *S. cinerearia* have been  
261 causing significant disruption to the urban forests, although some chemical control focused on the host.  
262 Similarly, *Simulium chutteri* broke out periodically in South Africa, even though some integrated  
263 control measures have been adopted (Rivers-Moore et al., 2008) . The issues remained as neither  
264 increasing the diversity of urban forest nor host-specific pest management measures may completely  
265 solve periodic pest outbreaks, which could be explained by the theory of island biogeography and  
266 landscape ecology (Simberloff, 1974; Zimmerman and Bierregaard, 1986). Our study focused on an  
267 area of the urban ecosystem at the local scale, where the urban trees were always isolated by some  
268 artificial surfaces and infrastructures, such as road networks and buildings (Wanghe et al., 2019).  
269 Consequently, each isolated land unit of urban greenery (e.g., an urban park) could be regarded as an  
270 analogous island (Zimmerman and Bierregaard, 1986). Because of the homogenous environment at the  
271 local scale, these widely planted urban greening trees (Figure 5) in each island likely occupied similar  
272 niches (Shushan et al., 2016), whose community structure would adapt itself to a dynamic equilibrium  
273 among the parasitoid and host, as well as their habitat (Zimmerman and Bierregaard, 1986).

274 Our results highlighted that the targeted insect's connectivity and migration should be considered a  
275 significant component within pest management. In many cases (Dodds and Orwig, 2011; Haack et al.,  
276 2010; Paap et al., 2017; Santini et al., 2013; Smith et al., 2007), some clear evidence proposed that  
277 after the arrival of these pests into urban areas, their subsequent spread into natural or planted forest  
278 landscapes was through multiple means and migratory pathways. For example, the invasive pathways  
279 of Asian long-horned beetles contained more than 15 families of trees, which remained in the urban  
280 settings until the 2008 outbreak in the natural forest in Massachusetts, the United States (Dodds and  
281 Orwig, 2011; Haack et al., 2010). In the current study, although *S. japonicum* was the primary host and  
282 provided the most suitable habitat to *S. cinerearia*, other tree species might still provide multiple niches  
283 to the moth (Zhang et al., 2013). The current study identified 42.9% of the occurrence points near but  
284 not located in *S. japonicum*. This result indicated that the imago of *S. cinerearia* could migrate and  
285 select multiple niches (tree species) other than its primary host (Figure 5). In addition, the result of the  
286 connectivity model (Figure 6) identified that not only the host tree species of *S. cinerearia* helped the  
287 pest to migrate, but other tree species also served the same purpose. We built and tested different  
288 scenario comparisons, suggesting that there would be multiple migratory pathways of *S. cinerearia* if  
289 the most suitable habitat (i.e., the primary host) had been removed. Those representations possibly  
290 supported the hypothesis of insects' multiple niches because of their migration and adaptation abilities.

291 We presumed that decreasing the pests' connectivity at some momentous pathways (e.g., the pinch  
292 points in Figure 6) might sufficiently alleviate the probability of periodic outbreaks and consequential  
293 spread. *S. japonicum* was the most vulnerable to the incursion of *S. cinerearia*, compared with the other  
294 widely planted nine tree species found in the area. A study suggested that *S. japonicum* was not the  
295 only food source of *S. cinerearia* (Tang et al., 2014). The pest prevalence in other plant species  
296 indicated that it could survive at different habitats in an urban ecosystem. Therefore, spraying  
297 pesticides only in its favorable habitat (*S. japonicum*) would not eliminate the outbreak, as some



298 survivors with <sup>33</sup>istance genes could move to the other niches through some high connectivity  
299 corridors (Feng et al., 2019; Tang et al., 2019). As a result, there would always be a risk of a secondary  
300 pest outbreak. Those representations might explain the reasons for periodic pest outbreaks in our study  
301 area, even if some chemical control had been conducted on the most vulnerable tree (i.e., *S. japonicum*).  
302 The other habitat providing critical landscape connectivity, such as sub-part B20 (Figure S2a), was  
303 also the potential and efficient pest management region (Figure 9). Our results showed that the genus  
304 *Populus* (*P. tomentosa* and *P. canadensis*) presented the most miniature habitat suitability. In other  
305 words, it was more resistant to *S. cinerearia* than the other eight species. Therefore, planting *P.*  
306 *tomentosa* and *P. canadensis* at “pinch points” would probably halt the outbreak and migration of *S.*  
307 *cinerearia*. Moreover, some anthropogenic barriers should be built in the “pinch points” to decrease  
308 the connectivity. For example, although the Sixth Ring Road isolated the Populations B and C (Figure  
309 6), some “pinch points” between the two populations still acted as corridors for *S. cinerearia*. Thus,  
310 cutting off those “pinch points” could minimize the probability of migration between Populations B  
311 and C.

#### 312 **4.2 The availability and limitation of this combinational model for pest management**

<sup>22</sup>  
313 Many studies (Fischbein et al., 2019; Jiang et al., 2018; Müller et al., 2019) recognized potential habitat  
314 aimed at pest management by SDM without analyzing <sup>2</sup>andscape connectivity. Thus, some significant  
315 pathways of the target species could not be identified (Cowley et al., 2015). For example, <sup>the</sup> potential  
316 <sup>dispersal of the</sup> Asian long-horned beetle was predicted using (but not combining) SDM and the circuit  
317 theory (Elmes et al., 2019), but the relationship between habitat suitability and landscape connectivity  
318 was ignored. However, few studies used this associated method on pest insect management (Elmes et  
319 al., 2019). Nonetheless, combining these two methods could identify a more practical pest management  
320 strategy without increasing the fieldwork labor and data input. In the current study, we illustrated a  
321 straightforward working procedure of this <sup>15</sup>mbinational modeling (Figures 2 and 9). This integrated  
322 method has also been empirically applied <sup>to predict the</sup> expansion <sup>of species</sup> invasion (Elmes et al.,  
323 2019; Falaschi et al., 2018), distribution range shift of endangered mammals (Puddu and Maiorano,  
324 2016; Zhao et al., 2019), the process of gene flow of a disease vector (Medley et al., 2015), and  
325 phylogeographic analysis (Dupas et al., 2014). These case studies identified that this combination  
326 model was available to different ecological processes.

<sup>12</sup>  
327 Geographic information system (GIS) has grown in popula<sup>12</sup> in recent years, and it has realized the  
328 integration of various types of data and algorithms, providing sophisticated tools for analyzing, editing,  
329 <sup>mapping, and performing spatial queries on data</sup> (Günen, 2021a). Meanwhile, some integrated GIS-  
330 based spatial models have improved the decision support processes in many research fields, such as  
331 energy planning (Günen, 2021b), urban greening planning (Pel<sup>18</sup>ssso et al., 2014), endangered animal  
332 conservation (Ahmad et al., 2020b, 2020a), and even helping <sup>to clarify the extent and impact of the</sup>  
333 COVID-19 <sup>pandemic</sup> (Franch-Pardo et al., 2020). In this paper, we proposed an integrated GIS-based  
334 modeling framework to improve research and applications focused on pest management. Three  
335 advantages were concluded as follows. Firstly, the researcher easily acquired the input data, only  
336 including the occurrence points (Figure 1) and eco-physiological environment layers (Table 1) of the  
337 targeted pest. Secondly, the results explicitly determined the relationship between the pest and the  
338 environment of its habitat (Table S3), as well as the potential regions for pest management. These  
339 results mapped the pests' vulnerable regions for forestry managers (e.g., the suitable habitat in Figure  
340 4 and the pinch points and corridors in Figure 6). Thirdly, the management effectiveness (Figures 7  
341 and 8) of different scenarios were reasonably quantified. Thus, customized solutions could be proposed  
342 in various pest situations (scenario B20 in Figure 9).

343 We concluded that our model has great potential in designing pest management policies and  
344 strategies in other similar situations around the world. Nonetheless, this approach had some limitations,  
345 which could be eliminated in further studies. For example, the spatial autocorrelation and pseudo-  
346 replicates may decrease the results' confidence (e.g., AUC value). Because our occurrence points were  
347 only obtained by our fieldwork at the local scale, whose spatial autocorrelation and pseudo-replicates  
348 were likely ignorable (Guisan et al., 2017), we did not examine the spatial autocorrelation in model  
349 residuals. Instead, if the occurrence points were from different sources at a large scale, the spatial  
350 autocorrelation should be detected by the Moran's I test on the generalized linear models (Guisan et  
351 al., 2017). Additionally, our results lacked validation. In a future study, we would experimentally  
352 implement scenario B20 to validate this model.

353

## 354 5 Conclusions

355 Tree being damaged by pest is global problem, especially in metropolitan cities. To eradicate pest and  
356 manage its outbreak, it is necessary to identify suitable habitat, corridors, and cut off the migratory  
357 pathway rather than only insecticide application. Because the only chemical eradication is not only  
358 expensive<sup>13</sup> laborious but also disturb the biological control mechanism and alleviate ecosystem  
359 function. Landscape connectivity has been defined as the degree to which the landscape facilitates or  
360 impedes movement among different patches, is a critical component of ensuring healthy ecosystems.

361 In the current study, we proposed a spatially explicitly GIS-based modeling framework, including  
362 MaxEnt and circuit theory to identify optimal regions for insecticide application. The AUC value  
363 indicated the parsimony results of SDM. The high AUC value indicated that the quality of habitat  
364 suitability map (derived by the MaxEnt models) was high and the sampling quadrats were sufficient.

365 This modelling framework can minimize damage to the natural ecosystem related to the unchecked  
366 and excessive use of pesticides and decrease the risk of periodic outbreaks. These optimal locations  
367 were targeted on potentially suitable habitat and the critical migration pathways, supporting precise  
368 insecticide application to save operation costs (e.g., travel and human efforts). Furthermore, this  
369 proposed modeling framework could be rerun for different parameters and newly collected points  
370 data to apply to similar urban areas for pest management.

## 371 6 Conflict of Interest

372 The research was conducted in the absence of any commercial or financial relationships that could be  
373 construed as a potential conflict of interest.

374

## 375 7 Acknowledgments

376 This study was supported by the China Postdoctoral Science Foundation No. 2021M693373 and the  
377 Natural Science Foundation of Qinghai province [2020-ZJ-912] and National Natural Science  
378 Foundation of China under Grants [32071489]

## 379 8 Author Contributions

380 XG, WK, and KL contributed to conception and design of the study. XG and KZ investigation and  
381 sample collection. XG, SA and WK performed the statistical analysis and wrote the first draft of the

382 manuscript. IIS, GN, KL wrote sections of the manuscript. All authors contributed to manuscript  
383 revision, read, and approved the submitted version.

## 384 **9 Supplementary Material**

385 Table S1: The occurrences database of *Semiothisa cinerearia*, Table S2: The detail results of  
386 Pearson correlation analyses among the eight environmental variables, Table S3: The relative  
387 contributions of eco-physiological variables used in the MaxEnt model for habitat suitability of  
388 *Semiothisa cinerearia*, Figure S1: The area under the receiver operating characteristics curve (AUC)  
389 of the MaxEnt model, Figure S2. The pest management locations of Scenario Bn (red and grey  
390 patches). The grey patches are the locations of hosts (sub-part A described in section 2.7 of the main  
391 body). The red patches are the locations of the other trees (sub-part Bn described in section 2.7 of the  
392 main body). From scenarios B20 to B100, the pest management locations are enlarging based on the  
393 range of landscape connectivity ranked by Jenks natural breaks.

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632 **Table**

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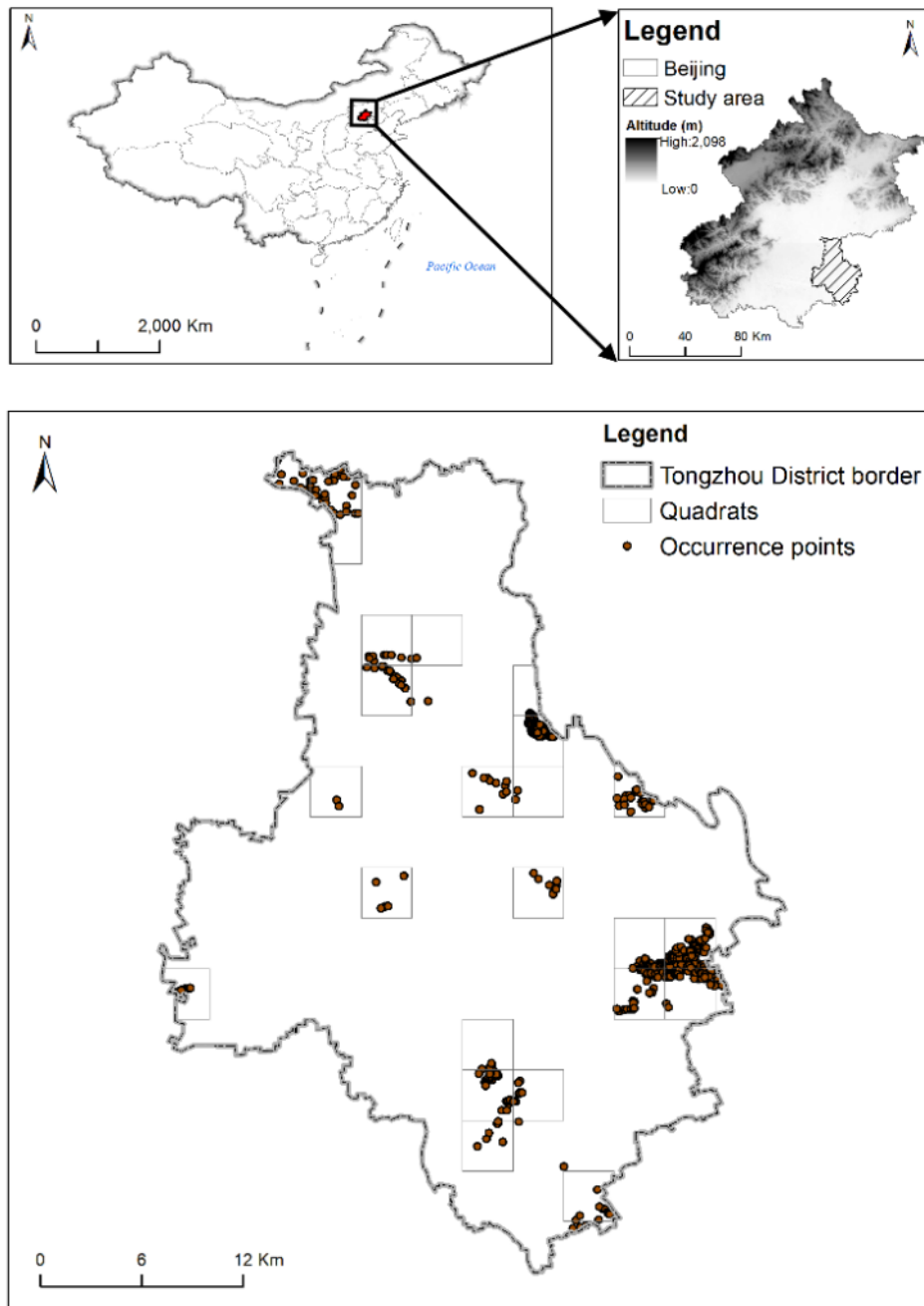
**Table 1.** List of the environmental variables.

Sources	Names	Descriptions	Ranges
The forest resource inventory database	Host	The distance to the nearest tree <i>Styphnolobium japonicum</i> (the primary host of <i>S. cinerearia</i> )	0 – 2,084 m
	Forest	The percentage coverage of forests in a raster	0 – 100%
	Shrub	The percentage coverage of shrubs in a raster	0 – 100%
	Meadow	The percentage coverage of meadows in a raster	0 – 100%
The land-use dataset	Building	The distance from a raster to the nearest buildings	0 – 1,086 m
	Road	The distance from a raster to the nearest roads	0 – 1,694 m
	Cropland	The distance from a raster to the nearest croplands	0 – 3,569 m
	Water	The distance from a raster to the nearest water	0 – 2,381 m

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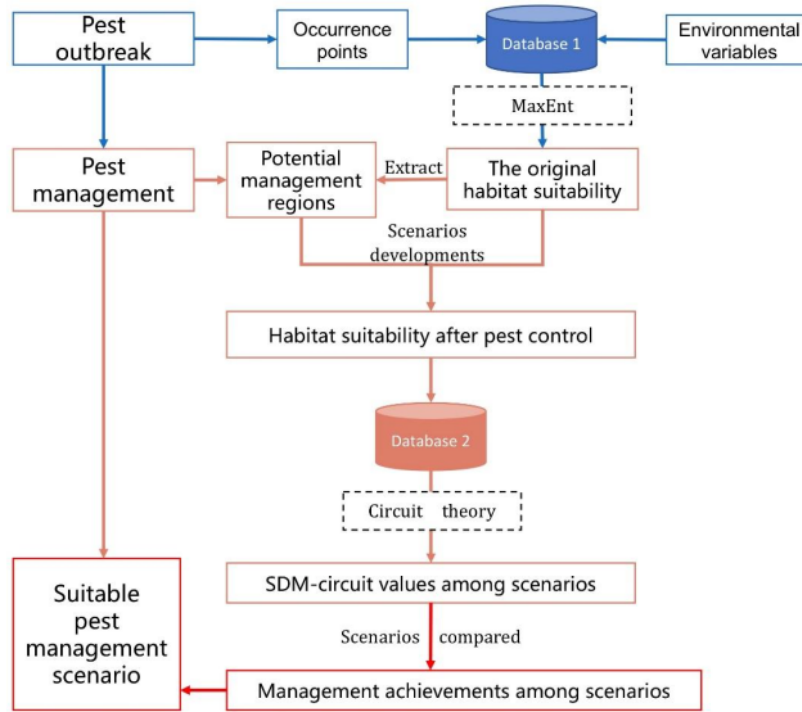
637 **Figures**



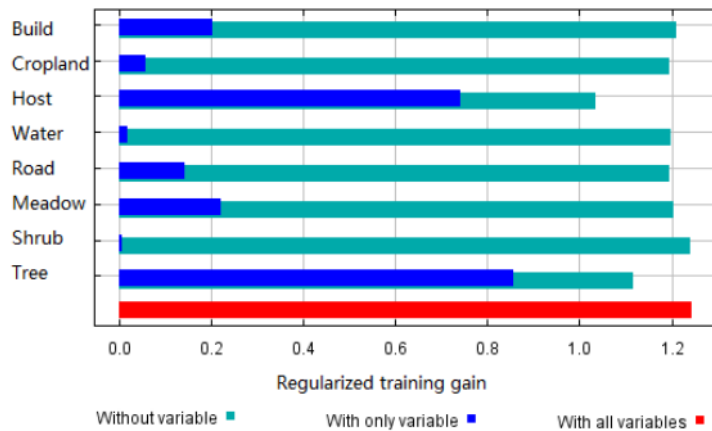
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640 **Figure 1.** Locations of our study area and the quadrats selected randomly, as well as *S. cinerearia*  
641 occurrence points surveyed.

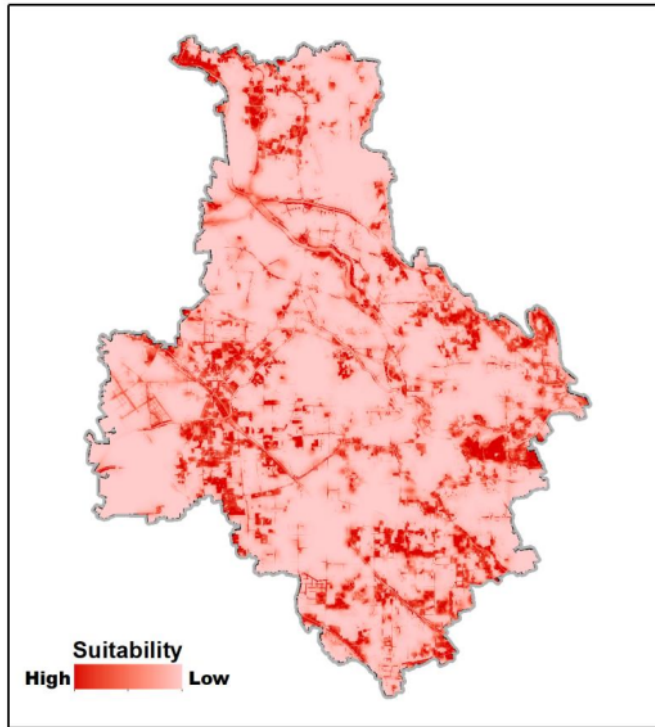


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644 **Figure 2.** Modeling framework of the study.



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646 **Figure 3.** Relative importance of environmental variables (training gains) in MaxEnt models by the  
647 Jack-knife test. The names and descriptions of those environmental variables see Table 1.

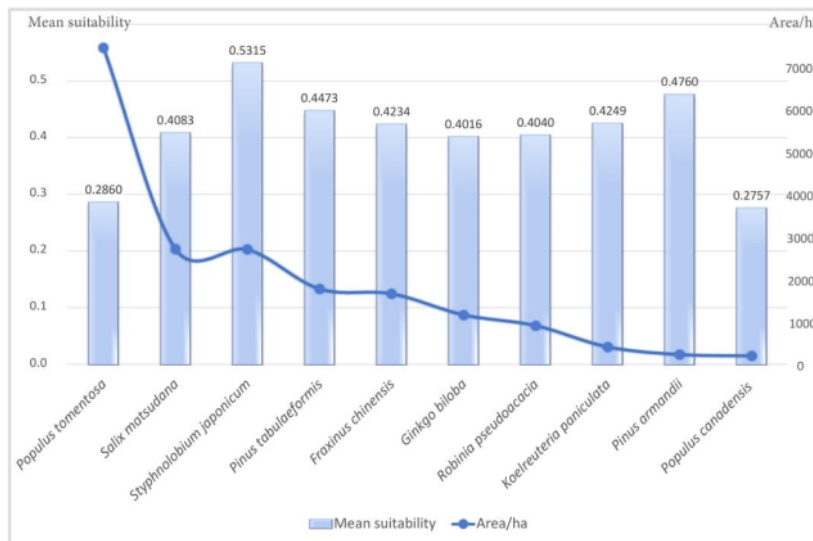




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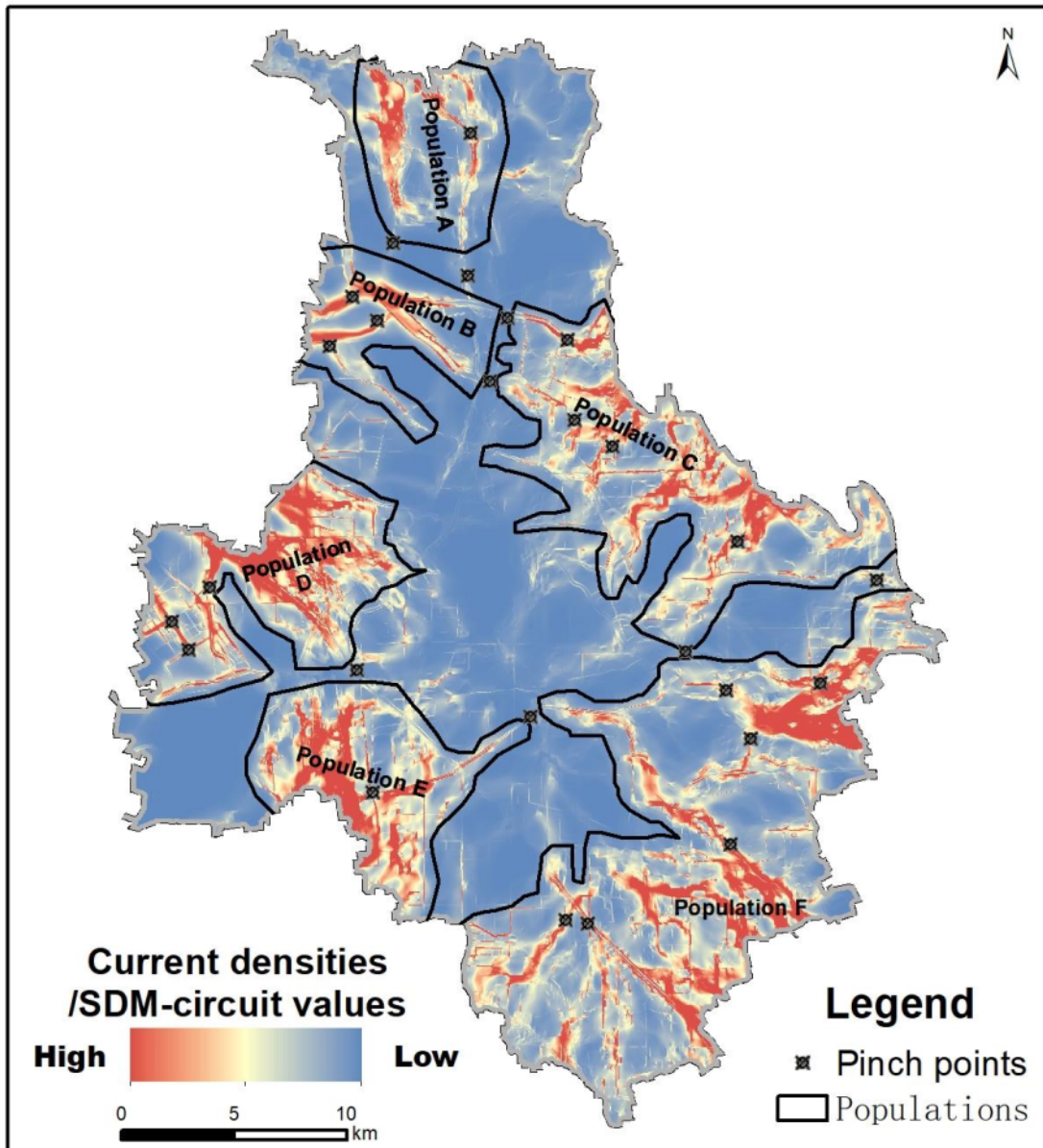
**Figure 4.** Habitat suitability map for *S. cinerearia*.



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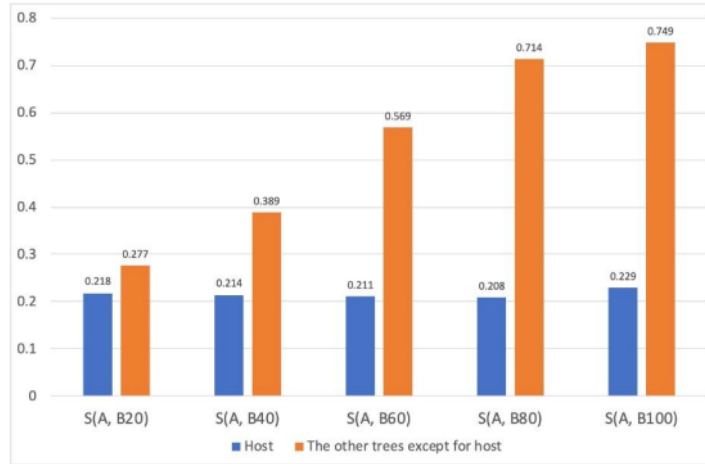
**Figure 5.** Comparison of habitat suitability providing to *S. cinerearia* among the top ten widely planned urban greening tree species.

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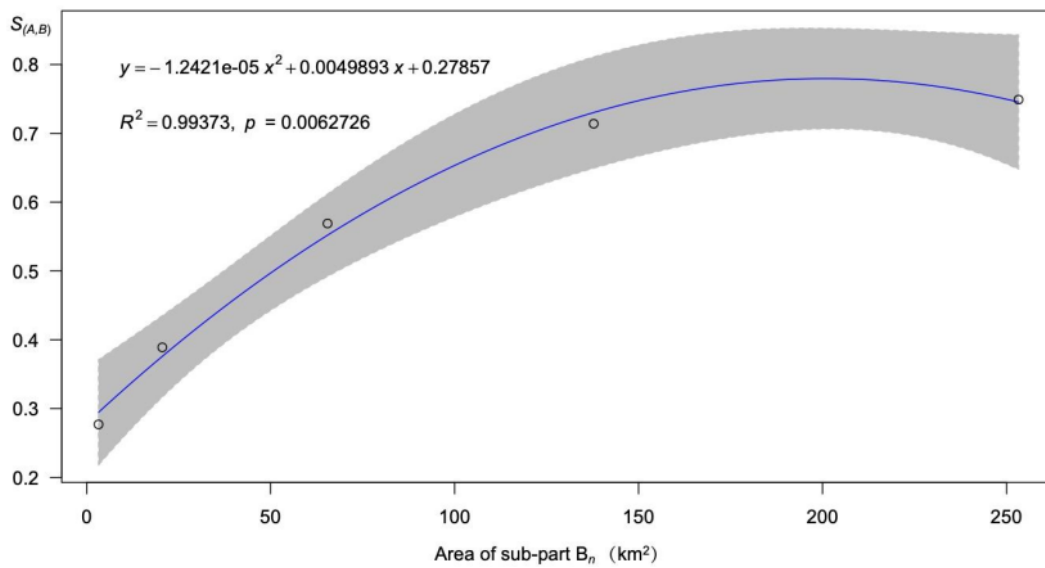
654

655 **Figure 6.** Map of the current densities/SDM-circuit values of *S. cinerearia*. Higher current densities  
 656 indicate cells with higher net passage probabilities for random walkers moving from one patch to the  
 657 other. The map highlights "pinch points", which are in some linear narrow pathways but providing  
 658 critical habitat connections between or within two populations.



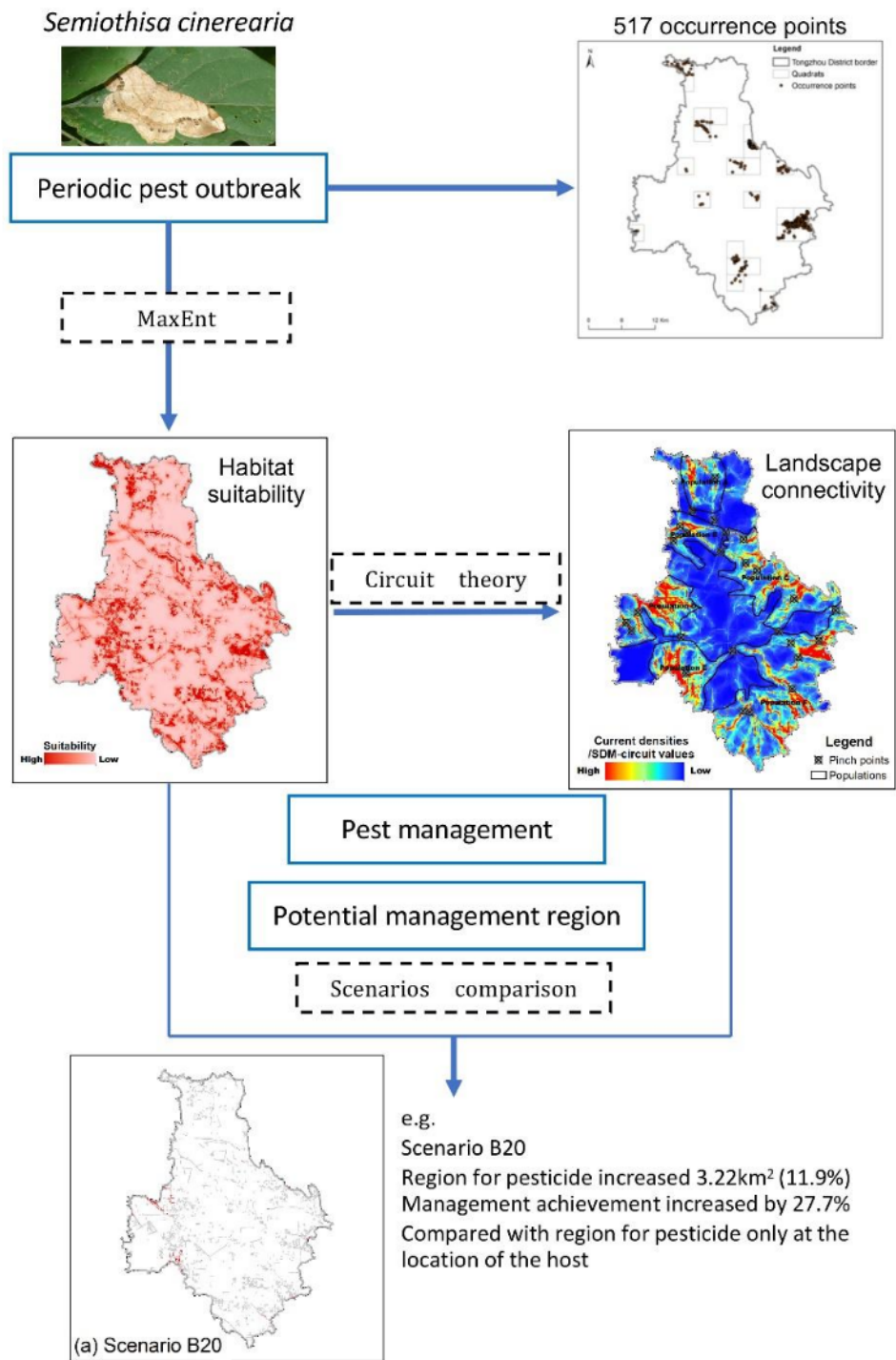
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660 **Figure 7.** The management performances among the scenarios. The y-axis is the value of  $S(A,B)$   
 661 described in Equation 2. In the x-axis, “A, Bn” means the scenario Bn vs. the original connectivity  
 662 model.



663

664 **Figure 8.** Result of the regression model. The y-axis 34 the value of  $S(A,B)$  described in Equation 2.  
 665 The x-axis means the area of sub-part Bn (km<sup>2</sup>). The grey shaded areas are 95% confidence intervals.



666

667 **Figure 9.** Flow chart for the combinational modeling between habitat suitability and landscape  
 668 connectivity to identify the potential regions for pest management.

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